

Chapter 4

The Distributive Effects of Biodiversity Policies: Dynamic Analysis

4.1. Intergenerational equity: evaluating costs and benefits across time

Biodiversity policies have an explicit time dimension. The total economic value of biodiversity concept contains some important intertemporal components, such as option values, exploratory values or quasi-option values (Bulte and Withagen, 2006), and bequest values (Pearce and Moran, 1994). These values are conceptually tied to the future in the following ways:

- Option values arise from the continued preservation of habitats and ecosystems by allowing conversion decisions to be postponed into the future.
- Quasi-option values arise from new information becoming available in the future that allows new decisions about biodiversity management to be made.
- Bequest values arise from the ability to pass on habitats and ecosystems to future generations.

The time scales over which these value components become relevant range from the very short to the very long. Progress in the life sciences may very rapidly allow the identification of valuable genetic traits in plants or micro-organisms that cannot currently be identified. The time scale of the exploratory or quasi-option value may therefore be measured in months or a few years. On the other hand, in the context of climate change, society may want to postpone decisions to irrevocably convert habitats given that it is not clear at the moment to what extent habitats may be degraded as a result of changing precipitation patterns. Here the time scale would appropriately be measured in decades or rather hundreds of years.

Given this explicit time dimension, policy decisions today will affect individuals currently alive, as well as generations not yet born. The policy-making process therefore needs to compare benefits and costs of biodiversity policies that may arise at vastly different points in time and justify them against some measure of intergenerational equity. Methods to do this are commonly referred to as discounting techniques.

Discounting is a major concern for intergenerational equity in biodiversity policies: the longer the time horizon of the effects of a specific policy, the larger the impact of discounting. Hence, the evaluation of policies involving irreversible components, such as species extinction, loss of habitats and ecosystems depends

to a large extent on the choice of the discounting model and its parameters. While this book can only broadly cover the issue of discounting, more detailed but accessible treatments (though not specifically on biodiversity policies) can be found in Pearce *et al.* (2003) and Groom *et al.* (2005).

4.2. Discounting

The economic tool for project evaluation is cost-benefit analysis (CBA). The basic rule is that if the social benefits exceed the social costs of a particular policy then it increases social welfare and should be implemented. This is a straightforward concept if both costs and benefits occur at the same instant or at least within a reasonably short time period, *e.g.* logging a single tree to obtain firewood in an otherwise intact forest. The costs of logging and the benefits of consumption occur in close succession.

However, if there is a considerable time interval between the two, for example if the costs have to be incurred right away while the benefits occur at some stage in the future (an investment), then how do we compare flows at such different points in time? Box 4.1 explains discounting. Extending the above example, if the tree is felled to build a house rather than to heat it, then the benefits occur over a longer period, *i.e.* the lifespan of the house. On the other hand, if the forest from which the tree is taken is close to collapse, the logging might have a significant impact on the survival of this ecosystem and the species therein. Hence, the costs are long-term too and might even contain irreversible elements, *e.g.* if some species are unique to this forest.

Box 4.1. Discount factors

Discounting is a method that systematically assigns different weights, called *discount factors*, to costs or benefits occurring at different points in time. These weights decrease over time, rendering distant costs and benefits less important. The conventional form of discounting, called “exponential discounting” uses a constant *discount rate* (s) to calculate discount factors w_t . The appropriate formula is then:

$$w_t = \frac{1}{(1+s)^t}$$

In principle, it is clear how individuals deal with the problem of assigning weights to future flows (*e.g.* payments) and there is a sizeable theoretical and empirical literature on how individuals discount future payoffs (Frederick *et al.*, 2002). However, it is very much debated how society as a whole should value costs and benefits occurring at different points in time and to different

generations. Individuals usually prefer benefits now to benefits in the future, and benefits enjoyed by their children to benefits enjoyed by their great-grandchildren. This applies to money as well as to risks (Pearce *et al.*, 2003). Hence, if people's preferences count in policy evaluation, this impatience should show up in the cost-benefit rule.

In contrast to this, philosophers and prominent economists have argued in favour of a zero social discount rate (Broome, 1992; Ramsey, 1928; Solow, 1974) whereby the current generation should receive the same weight as all generations to come. One reason for discounting usually acknowledged by this school of thought is the fact that in each period there is a positive but very small probability that the human race will become extinct (Stern, 2006), perhaps by the impact of a meteorite or a highly infectious and deadly disease for which no antidote is found. Hence, there is a chance that future generations might not exist and hence any cost or benefit occurring to them can be discounted accordingly.

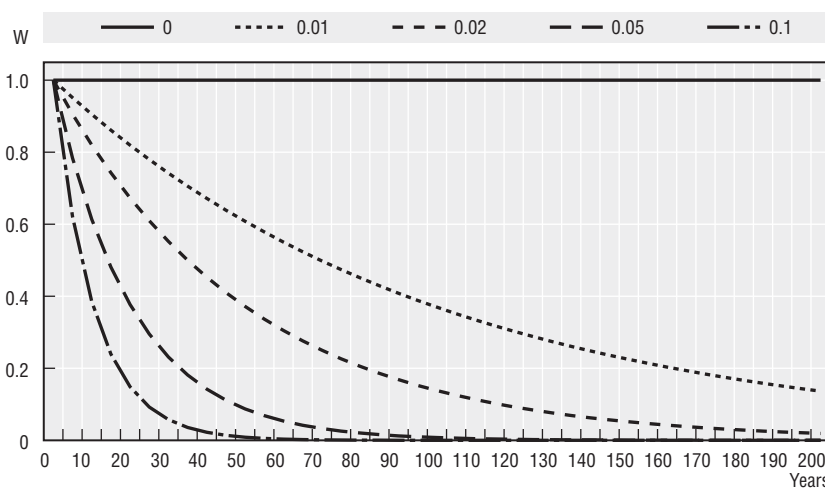
But even if present and future generations are treated equally, a further reason for discounting consumption (rather than utility) is that future generations are likely to be better off than current ones. Hence, given a decreasing marginal utility of consumption, an additional unit is worth less according to the future generation's own preferences than to the current ones. This effect competes with a contrasting one running in the opposite direction: there are some goods, such as many environmental amenities and biodiversity used for bioprospecting, whose availability does not increase at the same speed as consumption of manmade goods and for which no close substitutes are available. The marginal utility derived from such goods increases over time. This effect is reinforced if the supply of such goods declines due to conversion of natural landscapes, biodiversity loss and environmental degradation (Krutilla, 1967).

The probability that the human species will become extinct in any period is (by orders of magnitude) smaller than discount rates deduced from individuals' behaviour (see Frederick *et al.*, 2002). Moreover, in the latter there is a gap between developed and developing countries. While for the former, discount rates below 10% are common, for the latter, values above 25% and even above 100% have been estimated (GEF, 2006), reflecting the specific planning conditions in developing countries such as lower life expectancy, liquidity constraints and lower security of property rights. Hence, if policy choice is to be based on individual preferences, it might be crucial whether future costs and benefits occur to individuals living in developed or in developing countries. Moreover, applying discount rates based on empirical evidence in developed countries (*e.g.* 3.5% for the UK, see HM Treasury, 2003) can result in a serious lack of acceptance by local stakeholders in developing countries. If the benefits of the policy occur in the future, stakeholders might

put a considerably lower value on them than the planner assumes. This is an important constraint for the design of voluntary and incentive-based biodiversity policies.

How the choice of the discount rate matters in the medium and long run is illustrated by Figure 4.1. It presents the evolution of discount factors corresponding to different discount rates. Discount rates of 0, 0.01 (1%), 0.02 (2%), 0.05 (5%) and 0.1 (10%) per year are shown over a 200 year period. The lines show what happens to an initial quantity ($w = 1$ in year 0) as a result of those discount factors. So, for people with a discount factor of 2%, a promise of EUR 100 in 40 years is today worth only EUR 45 ($w = 0.45$ in year 100). Alternatively, if a foreseen event was to cause a loss of EUR 100 in 40 years, then we would only be willing to pay EUR 45 to avoid that loss. In other words, a predictable loss of EUR 100 40 years from now might go unmitigated.

Figure 4.1. **The evolution of the discount factor over time for different constant discount rates**



What is the effect of discounting on decisions? By attaching lower values to costs and benefits occurring in the distant future, discounting (and hence, the choice of s) has a major impact on the outcome of cost-benefit analysis and project evaluation. This is especially so for distributive effects between generations where, by definition, long time horizons are involved and thus discounting is a key determinant for identifying a desirable policy. Higher discount rates imply lower importance attached to future generations. To illustrate this point, consider the examples in Table 4.1.

A biodiversity conservation project costs 1 million right now but yields 5 million in conservation benefits 50 years in the future. Whether the project passes the cost-benefit test depends on the discount rate chosen (see Table 4.1). For $s = 2\%$, the discount factor for $t = 50$ is 0.3715. Hence, the net benefit is $NB = 0.3715 \times 5 \text{ m} - 1 \text{ m} = 0.8575 \text{ m}$ and is thus greater than zero. The project is beneficial at a social discount rate of 2%. Repeating the same exercise with $s = 5\%$ yields a very different result. The corresponding discount factor is 0.0872 ($w_{50}[s=5\%]$). The net benefit is negative ($NB = 0.0872 \times 5 \text{ m} - 1 \text{ m} = -0.564 \text{ m}$). Hence, at a 5% social discount rate, the project would not go ahead.

Table 4.1. **Two hypothetical cost-benefit scenarios with exponential discounting**

Costs		Benefits		Discount rate s	Discount factor w	Net benefit	Evaluation
Amount	Year	Amount	Year				
1 million	0	5 million	50	2%	0.3715	0.8575 million	Desirable
1 million	0	5 million	50	5%	0.0872	-0.564 million	Not desirable

Discounting therefore affects the set of socially desirable policies, as well as putting constraints on their implementation. Careful consideration of discounting and its effects are therefore key for successful biodiversity policies.

4.2.1. The problem of discounting

Exponential discounting at a positive rate has been attacked as a “tyranny of the present”. If very long-term policy decisions are considered, such as conversion of pristine land, flooding due to dam construction or biodiversity loss, any costs or benefits occurring to future generations receive little to almost no consideration in current decisions (see Figure 4.1). Hence, although distributive effects of some biodiversity policies might be huge, they would frequently be dwarfed by the application of discounting.

The immediate relevance of the discount rate to biodiversity policies is a mainstay of natural resource economics: Clark (1973) demonstrated that high intertemporal discount rates are a key reason why many managed species have been “rationally” driven to, or close to, extinction in the past, because the future benefits of their existence have been considered negligible when decisions were taken. Swanson (1994) extended Clark’s (1973) framework to species and habitats that are not managed, and showed that the same logic can also explain habitat conversion and deleterious management practices that give rise to “extinction by neglect”. Discounting and sustainable development are often regarded as irreconcilable.

The most popular proposal for solving the “tyranny of the present” is to abandon discounting altogether (i.e. using a zero discount rate). This

essentially gives the same weight to all current and future generations, including those living a million years from now. Hence, any project that at some stage yields benefits that are greater than the costs (both undiscounted) is worthwhile. However, zero discounting, taken seriously, has important implications for broader macroeconomic decisions such as the savings rate. Savings should by far exceed their current level and consumption by the current generation should fall considerably in order to yield high benefits to a generation in a far distant future. In fact, current consumption might conceivably be at risk of being driven down to subsistence levels. On a more abstract level, zero discount rates also raise the possibility that it is no longer possible to even formulate an optimal consumption and savings path (Koopmans, 1965; Asheim et al., 2001). Put somewhat more pointedly, using a zero discount rate might have prevented mankind from converting any pristine land and from using any non-renewable resources. Zero discounting has therefore been labelled “tyranny of the future”.

4.2.2. Intergenerational equity: the role of uncertainty and risk

In terms of intergenerational equity, both a constant positive discount rate and a constant zero discount rate lead to unsatisfactory policy outcomes over the long time scales that characterise many biodiversity policies. So does a balanced solution exist?

One hopeful candidate is “hyperbolic discounting” (Box 4.2). While using strictly positive discount rates, it differs from exponential discounting in one important respect. The discount rate s is not constant but decreases over time. Hence, discount factors decrease less than they would for constant discount rates in the long-run.

One major argument in favour of declining discount rates is uncertainty about future states of the world. Two conceptualisations of this uncertainty have been proposed. While Weitzman (1998) assumes uncertainty over the future discount factor, Gollier (2002a, b) allows for uncertainty over consumption paths. Both approaches come to the same conclusion: discount rates are declining. Uncertainty over future states of the world is common in biodiversity policies. The bioprospecting value of species, the effect of losing ecosystem services and the preferences of future generations are all highly uncertain.

Declining discount rates are backed by a very different recent theoretical approach in the social choice tradition. If a social planner advocates a mixed goal which combines a high discount rate and a low (zero) discount rate, the result is a social discount rate that declines over time (Chichilinsky, 1996; Li and Löfgren, 2000). Moreover, hyperbolic discounting is supported by empirical evidence (see Frederick et al., 2002); people seem to apply hyperbolic discounting in their everyday decision-making.

However, there is a drawback to the concept of declining discount rates. Most types suffer from what is called time-inconsistency. Using varying discount rates current optimal plans might not be consistent with what the same individual regards as optimal in the future (even in the absence of uncertainty about future states of the world and preferences). Hence, one does not stick to the original plan, *i.e.* policies are revised (if possible) as time passes and these deviations are (or at least could be) anticipated. If policy outcomes are irreversible, there might be regret. For example, although it seemed optimal to a social planner to convert some parcel of land, he/she later might regret this decision but be unable to restore its original state. While the occurrence of time-inconsistency under time-varying discount rates is widely accepted, there is a debate about whether this is actually a relevant problem (Pearce *et al.*, 2003).

Throughout, a key challenge in comparing costs and benefits of biodiversity policies across time by means of discounting is to answer the question of what constitutes the “right” discount rate. While this question arises with exponential as well as with hyperbolic discounting, with the latter it has an additional dimension. The problem is not only to pick a single parameter (which is difficult enough) but to choose an entire profile of discount rates. The use of hyperbolic discounting in the UK (Box 4.2) shows that important recent advances in addressing this problem can be reasonably implemented in real world policy. Transferring the approach to biodiversity policy would be straightforward.

4.3. Heterogeneous generations

Differences in discount rates used by individuals or groups within a generation are perhaps even more common and exaggerated than across generations. Box 4.2 shows the UK’s discount rate starting at 3.5% in the early years, but then falling over time. The decline in the rate for UK policy is considerably smaller than the difference in discount rates across countries (Table 4.3). Rates vary by a factor of four, even within this relatively homogeneous sample. Less developed countries usually have higher discount factors as they rely on natural resources more than developed countries. Resource dependence is often associated with a higher discounting of future benefits since resources provide limited ability to distribute consumption over time. Poor countries that rely on consumptive aspects of biodiversity are more likely to manage their resources well, given that they rely on them for survival (though Diamond, 2005, provides some counter-examples). Nonetheless, subsistence is often used as a model of situations of high discount rates because the lack of sufficient reserves causes decisions to be made mainly on the basis of short-term considerations. Moreover, the non-consumptive

Box 4.2. Hyperbolic discounting in the UK Green Book

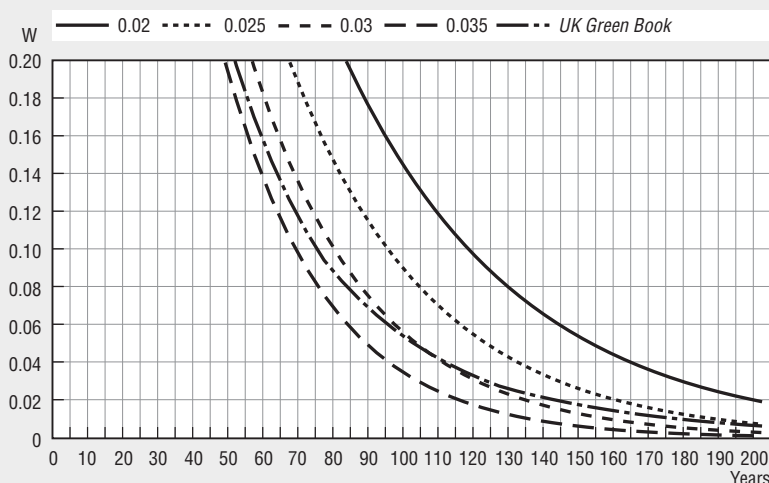
The evaluation of public policies in the UK is based on HM Treasury's (2003) *Green Book, Appraisal and Evaluation in Central Government*. For all projects with impacts lasting for less than 30 years a constant discount rate of 3.5% has to be used, based on empirical estimates for the UK. However, for policies with long-term effects the following pattern of discount rates is applied.

Table 4.2. **The declining long-term discount rate**

Period of years	Discount rate
0-30	3.5%
31-75	3.0%
76-125	2.5%
126-200	2.0%
201-300	1.5%
301+	1.0%

The effect of this stepwise decline in the discount rate is presented in Figure 4.2. While for the first 30 years the evolution of discount factors matches that of a constant 3.5% rate, for later periods the weight of future flows is significantly above that reference scenario, *e.g.* the weight put on any cost or benefit in year 200 is about six times as high with the declining discount rate than in the scenario with a constant rate of 3.5%; in year 300 the difference is already two orders of magnitude.

Figure 4.2. **Discount factors with decreasing discount rate of the UK Green Book**



Source: HM Treasury (2003), Annex 6.

Table 4.3. **Discount rates as listed by Commissariat Général du Plan in France**

Country	Discount rate	Time horizon (years)
South Africa	8%	20-40
Germany	3%	Variable
Australia	6-7%	20-30
Canada	5-10%	20-50
Denmark	6-7%	30
United States	3-7%	Variable
Italy	5%	
France	8%	30
Hungary	6%	30
Japan	4%	40
Mexico	12%	30
Norway	5%	25
New Zealand	10%	25
Netherlands	4%	30
Portugal	3%	20-30
Czech Republic	7%	20-30
United Kingdom	3.5%	30
Sweden	4%	15-60
European Commission	5%	
World Bank	10-12%	

Source: Hepburn (2006).

benefits of biodiversity are likely to be highly discounted since they are associated with leisure time, a limited commodity at low income levels.

Underpinning the notion that developing countries have higher discount rates than developed countries is the observation that “liquidity constraints” force individuals to behave as if they cannot plan for the long term. That is, even when they know that postponing an action (*e.g.* consumption) to the future will bring greater overall benefit, they may be prevented from acting on that knowledge when they are unable to borrow against future benefits. The classic example is the farmer whose circumstances force him/her to eat the seed crop that was to be used to plant next season’s crop. Such activity implies an extraordinarily high discounting of the future.

This heterogeneity across countries implies that when biodiversity is unevenly distributed globally, there will be differences in how much conservation one country is willing/able to undertake, and how much other countries would like it to undertake.

4.3.1. Intergenerational equity and intragenerational equity

As most biodiversity policies have both long-term effects and affect people of different wealth we need to trade-off distributional effects against those across generations. For discussion on how to incorporate intragenerational equity into cost-benefit analysis see Section 2.3 of this book.

Helping the poor of today might harm future generations or *vice versa*. Hence, it is important to choose a consistent way to treat people living at different points in time and with different economic status. If the interests of the poor today are valued more than the interests of the rich, and if future generations tend to be better off than current ones, there is a case to be made for applying the same principle in both situations. The *Stern Review on the Economics of Climate Change* (Stern, 2006) has been criticised for being inconsistent on this important issue (*The Economist*, 2006).

The main concept that links the two issues is the elasticity of the marginal utility of consumption. This states by how many percent the utility of a person increases if his/her income increases by one percent. This is thought to be roughly constant across income classes. Hence, giving the same amount of money to a poor person (for whom it adds significantly to current consumption) creates more utility than if it is given to a better off person (for whom it is just a further drop in an already large pool). For consistency of the two dimensions of equity the same income elasticity has to be used when calculating the effective discount rate and distributional weights used to adjust for different income levels among members of the same generation. The *UK Green Book* (HM Treasury, 2003) assumes an elasticity of one in both cases.

4.4. Summary and conclusions

Part I has introduced the key concepts relevant to the analysis of distributive impacts of biodiversity policies – efficiency, cost effectiveness and distributive impacts – and how they relate to policies for maintaining and improving biologically diverse habitats and ecosystems. It has explained the role of CBA for biodiversity policy-making and how the integration of efficiency rules based on CBA has strengthened the case for biodiversity policies to be considered as important alongside other policy issues. At the same time, it has stressed that by using the concept of net social gains, CBA is severing the ties with richer welfare-theoretic dimensions.

We have explained how distributive impacts can be empirically measured, quantified in terms of summary values, and communicated in a policy-making context. We have also presented a positive analysis of the distributive effects of biodiversity policies, drawing from a rich set of case studies and examples to explain the links between policy objectives,

instrument choice and welfare outcomes, while bearing in mind the success of the policy in maintaining and improving species-rich habitats and ecosystems. Important distributive dimensions that are dealt with are the spatial and the intertemporal distribution of welfare impacts associated with biodiversity-related policies. Our main conclusions are as follows:

- There are many suitable approaches for measuring distributive impacts, with differing data requirements and ability to capture these impacts. However, not all measures are equally useful in different conservation contexts. Hence, there is a need to develop criteria for judging the information contained in different measures and their adequacy for specific contexts.
- The distributive impacts of biodiversity policies are real and in many cases non-marginal. The primary distributive effects of biodiversity policies are likely to be pro-rich for both theoretical and empirical reasons. The secondary distributive effects are determined by the choice of instruments, which can mitigate or amplify the primary distributive effects. A wide variety of instruments is available for mitigating and potentially reversing distributive effects.
- The trend towards market-based instruments in biodiversity policies is likely to ameliorate the negative impacts on the poor of traditional instruments, such as protected area policies. However, there is evidence that while market-based instruments do not hurt lower income households, higher income households tend to profit relatively more.
- A “spatial mismatch of costs and benefits” (Wells, 1992) has been identified for some biodiversity policies, with local people often bearing most of the costs and populations of far-away countries receiving most of the benefits. However, if handled well, protection for conservation can be pro-poor because protected areas spell out – often for the first time – the precise nature of use rights in an area. Local communities may gain from protected areas because outside competitors are excluded.
- At the international level, current distributional problems are likely to persist. Many of the difficulties in translating the international willingness to pay for biodiversity conservation into flows of funds to areas of high conservation importance remain unresolved. The prevailing internationally agreed sharing rules for gains from international co-operation on biodiversity conservation contribute to this outcome.
- There is a significant intergenerational distributive dimension of biodiversity policies, since biodiversity policy decisions today will affect individuals currently alive, as well as generations not yet born. Ensuring that decisions taken today do not affect future generations can be addressed by varying the discount of costs and benefits arising at different

points in time (hyperbolic discounting). At the same time, consistency between inter- and intragenerational equity is required.

With these key concepts, measurement methods, and empirical observations in mind, we now turn to the question of whether policy-makers should consider and address distributive issues *within* biodiversity policies – or whether these distributive issues should be left out of the picture when deciding between different courses of action.

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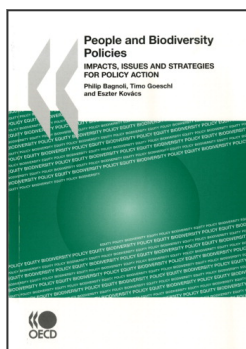
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